

Assessing the Synergistic Effects of Co-digestion of Maize Silage and Red Chicory Waste



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Nowadays, the anaerobic digestion of organic wastes to produce renewable energy is a reality in many countries. Many feedstocks can be processed by anaerobic digestion to produce biogas, however, anaerobic digestion of single substrates can have drawbacks that could be eliminated by the anaerobic co-digestion of their mixtures (two or more). In this paper, the anaerobic co-digestion of maize silage and red chicory (*radicchio*) waste is presented. Several batch anaerobic biodegradability tests were performed in order to compare the methane production of the blend with those of the two substrates digested separately. The methane production was modelled by a first order kinetic model, focusing on the initial substrate solubilization. The specific methane productions of maize silage and red chicory as single substrates were 0.346 L CH₄ g⁻¹ ODM and 0.326 L CH₄ g⁻¹ ODM. The first solubilization rate constants (k_{sol}) were 0.231 d⁻¹ for the digestion of maize silage, and 0.389 d⁻¹ for the *radicchio* waste. Two parameters representing the relative changes in specific methane production and solubilization rate were calculated in order to evaluate the synergistic effects due to co-digestion of studied substrates. Results showed that blending the two substrates enhanced the performances of the AD process, mainly with respect to biogas production kinetics with low increments in the ultimate methane potential.

Key words:

anaerobic co-digestion, anaerobic biodegradability, disintegration, specific methane production, theoretical methane potential

Introduction

The anaerobic digestion (AD) of organic waste is an attractive and sustainable technology for energy recovery and pollution prevention. The main product of AD is biogas (a mixture of CH₄ and CO₂) that may be used for heating and electricity production, and, if upgraded, may be used as vehicle fuel or injected in a natural gas network¹. A secondary product of the process is a sludge residue, called digestate, which can be directly used as soil amendment or as starting material for high-quality compost preparation².

Many feedstocks can be processed by AD to produce biogas, including the organic fraction of municipal solid waste (OFMSW), agro-industrial waste, agricultural residuals, energy crops, and sewage sludge. In recent years, Italy is witnessing a proliferation of biogas plants purpose-designed for energy recovery. The Italian biogas production is the second largest in the European Union after Germany³. According to the EBA Statistics⁴, at the end of 2015 there were more than 1500 biogas plants

operating in Italy. Most of these operate in co-digestion and, consequently, are fed with energy crops (mainly cereal silage), agricultural by-products (animal sewage), and agro-industrial residues⁵. It is recognized that the simultaneous digestion of two or more organic substrates offers several advantages, such as the improvement of the balance of nutrients and the C:N ratio, the alleviation of inhibitory effects due to toxic substances, and the enhancement of methane production kinetics^{6–11}. However, selecting the blend of substrates leading to a stable anaerobic co-digestion (AcoD) operation is not trivial, as it requires knowledge and expertise on the process¹². Co-substrate selection through optimization methods seems to be the way to facilitate synergistic effects during AcoD^{12,13}. As reported in literature^{14,15}, the best co-substrate blend should assure: (i) a balanced carbon-to-nitrogen (C:N) ratio; (ii) a micronutrients equilibrium; (iii) a dilution of inhibitory and toxic compounds; (iv) a higher readily biodegradable organic fraction; (v) an optimized methane production, and (vi) enhanced digestate stability. Usually, the main criterion adopted, among all, to decide the best ratio between waste for AcoD process, is the optimization of the C:N ratio¹⁵.

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In this paper, the effects of co-digestion of maize silage (MS), with waste coming from the harvesting of red *radicchio* (*Cichorium intybus*) (RR), were evaluated. The aim was to investigate the biovalorization via AcoD of RR waste by adding them to an ideal AD substrate, such as maize silage. Three scenarios were evaluated by means of anaerobic biodegradability tests: mono-digestion of maize silage and red *radicchio* separately, and co-digestion of mixed substrates. The methane production was modelled by a first order kinetic model, focusing on the initial substrate solubilization (disintegration plus hydrolysis). With a pragmatic approach, the synergistic effects due to substrate blending were assessed by calculating the relative change in the ultimate methane production and in the solubilization rate as a result of co-digestion in comparison with mono-digestion.

Materials and methods

Inoculum

Anaerobic sludge was collected from a biogas plant operating at a fruit and vegetable processing factory. Sludge was analyzed in order to determine dry matter (DM) and organic dry matter (ODM) according to Standard Methods¹⁶, while the total COD was measured according to the method developed by Raposo et al.¹⁷ Soluble COD was measured by Hach-Lange test cuvettes. Before performing the anaerobic biodegradability tests, the sludge was pre-incubated for 15 days in order to deplete any residual biodegradable materials. The endogenous (*i.e.* with no external feed) specific methanogenic activity (SMA_{end}) recorded during the incubation was $0.047 \text{ g COD-CH}_4 \text{ g}^{-1} \text{ ODM d}^{-1}$. After this degassing period, the sludge sample was diluted (almost 1.5 times) and fed into the reactor. The characteristics of the anaerobic sludge at the beginning of biodegradability tests were: 13.65 g L^{-1} of DM with a volatile content of 66.3 %; 12.48 g L^{-1} of total COD, and 1.56 g L^{-1} of soluble COD.

Substrates characteristics and calculation of the theoretical methane potential

The characterization of the organic substrates addressed to anaerobic digestion is the first step in order to provide an estimation of their biomethane potential, and, therefore, to evaluate the economics of the system. When dealing with lignocellulosic materials (as the substrates treated here), one must

pay attention to the content of lignin that is not degradable in anaerobic conditions, and that creates barriers for microbial degradation of cellulose¹⁸.

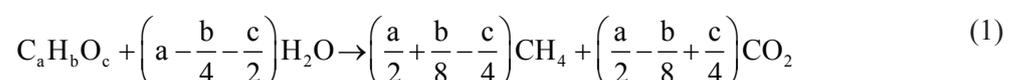
In the present paper, the chemical compositions of maize silage and red *radicchio* were adopted from literature.

The use of maize silage in biomethanation processes is well known^{19–22} and it is possible to state that, currently, biogas production is mainly based on the AD of this substrate⁵. Significant differences among the data can be noticed when comparing maize characterization from literature due to many factors, such as the plant variety, weather during cultivation, harvesting technology, and analytical methods²³. As an example, a maize characterization comparison is reported in Table 1.

In the present work, the characterization of Herrmann et al.²² was adopted, since their MS components values fall in an average range of Table 1.

Red *radicchio* of Treviso, named after the Italian region where it originated, belongs to an ancient crop cultivated in northern Italy (Veneto region)²⁴. The agronomic, economic, and social importance, and overall typicality of RR have been recognized by the European Union, which awarded it, at the end of the 1990s, with the PGI (Protected Geographical Indication) label²⁵. According to recent statistics²⁶, the Italian red *radicchio* production for the year 2016 exceeds 120,000 tons with an occupied area of approximately 7700 ha. Outside Italy, *radicchio* is now commonly being grown throughout Europe, Japan, the United States, Guatemala, Mexico and South America²⁷. The onerous manufacturing process, required by this chicory to achieve its high qualitative and aesthetic level, results in a great amount of waste, around 30 % of rejected *radicchio* heads²⁸, that can be subjected to anaerobic digestion. As regards the composition of *radicchio* waste, the initial fractionation of organic matter between carbohydrates, proteins, and lipids was obtained from the CREA (Consiglio per la Ricerca in agricoltura e l'analisi dell' Economia Agraria) tables²⁹ and, at a later stage, the total carbohydrates sharing was assumed from Wahid et al.³⁰

Several methods exist in literature to estimate the theoretical methane potential of the organic substrates starting from the content of carbohydrates, proteins and lipids, with particular attention to lignocellulosic components^{21,31,32}. In the present work, according to Angelidaki and Sanders³³, the theoretical methane potential of the substrates components was calculated by Buswell's formula³⁴:



which provides the specific theoretical methane potential, $B_{0,th ODM}$, related to volatile solids:

$$B_{0,th ODM} = \frac{\left(\frac{a}{2} + \frac{b}{8} - \frac{c}{4}\right) 22.4}{12a + b + 16c} \left[\text{STP} \frac{\text{LCH}_4}{\text{gODM}} \right] \quad (2)$$

Neglecting the contribution of lignin (not anaerobically biodegradable), the $B_{0,th ODM}$ of the complex substrates was then obtained as:

$$B_{0,th ODM} = \frac{0.415L \cdot x_{CH} + 0.496L \cdot x_p + 1.014L \cdot x_L}{ODM} \left[\text{STP} \frac{\text{LCH}_4}{\text{g ODM}} \right] \quad (3)$$

where x_{CH} , x_p and x_L represent the fraction of carbohydrates without lignin ($C_6H_{10}O_5$), proteins ($C_5H_7O_2N$), and lipids ($C_{57}H_{104}O_6$) of the complex substrates, respectively.

The characteristics of tested substrates with the calculated methane potential are reported in Table 2.

Experimental set-up

The anaerobic biodegradability tests were carried out in home-made equipment presented in Fig. 1. The anaerobic reactor consisted of a 5-L glass bottle placed in a controlled temperature environment (water bath with a Julabo MB heating immersion circulator) of 35 °C (± 0.1 °C), and mixed continuously with magnetic stirrers (ARE, Velp Scientific). A pressure transducer (RS Instrument) was connected to the bioreactor to outline the pressure changes during the test. A volumetric method with water displacement was used to measure the biogas produced³⁵, with composition achieved by a gas analyzer (GA 2000 plus, Geotechnical Instruments). All the data were finally recorded by PC. The assembled pipelines were made of stainless steel and PTFE.

Three scenarios were considered: mono-digestion of MS, mono-digestion of RR, and co-digestion of MS-RR. The methane potential of each typology of feed (MS, RR, and blend) was evaluated by means of consecutive batch tests with an applied substrate to inoculum (S/I) ratio variable from 5 % to 9 % in ODM basis. For a given feed, the production of biogas was monitored and recorded for a variable time interval (from one to three weeks)

Table 2 – Characteristics of tested substrates

Parameter	Units	Maize silage	Red radicchio	Notes
DM	[%]	33.3	5.3	(a)
ODM	[% DM]	95.8	88.2	(a)
Carbohydrates (*)	[% DM]	82.5	60.2	(b)
Lignin	[% DM]	2.9	5.9	(b)
Proteins	[% DM]	7.8	20.5	(b)
Lipids	[% DM]	2.6	1.5	(b)
Calculated COD	[gCOD g ⁻¹ DM]	1.21	1.14	(c)
C/N	–	35	13	(b–c)
COD/N	–	97	35	(b–c)
B_{0th}	[STP LCH ₄ g ⁻¹ ODM]	0.425	0.416	(c)

(*) without lignin

(a) from analysis, (b) from literature, (c) calculated

Table 1 – Literature data on characterization of maize silage

DM (%)	ODM (% DM)	Ash (% DM)	CP (% DM)	CL (% DM)	CF (% DM)	NDF (% DM)	ADF (% DM)	ADL (% DM)	NFE (% DM)	Starch (% DM)	Ref.
31.0	95.5	–	6.7	2.9	–	34.6	21.6	2.0	–	–	19*
30.2	95.8	–	7.8	2.6	–	41.2	24	2.9	64.7	–	22
29.5	81.4	5.5	8.9	3.1	19.2	43.9	23.9	2.3	–	22.1	47
31.1	93.2	6.8	10.3	5.1	15.5	71	33.4	11.6	62.4	–	36

DM: dry matter; ODM: organic dry matter; CP: crude protein; CL: crude lipids; CF: crude fibers; NDF: neutral detergent fibre; ADF: acid detergent fibre; ADL: acid detergent lignin; NFE: nitrogen free extract.

* data relative to maize after 180 days of ensiling

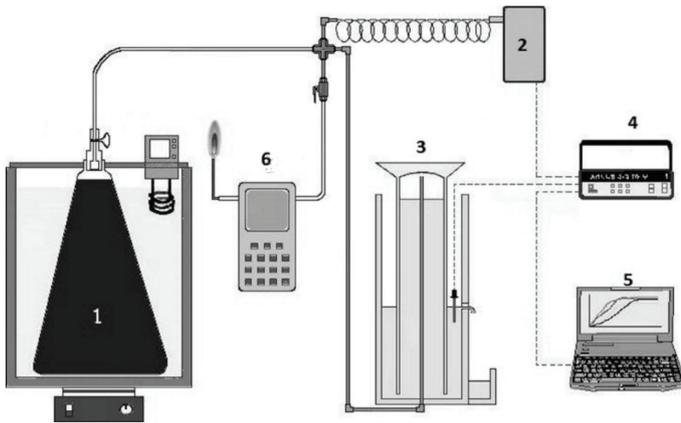


Fig. 1 – Experimental set-up: 1) anaerobic reactor; 2) pressure transducer; 3) gasometer; 4) data-logger; 5) personal computer; 6) gas analyzer

up to reaching the minimum biogas flow of 0.15 mL min^{-1} (value set as the proper limit for the production recording of the equipment).

The low S/I ratio values were selected in order to reduce the reaction times allowing to maintain a biomass with a good activity and to focus the initial solubilization of particulate organic matter^{36,37}. The strategy of operating consecutive batch tests was adopted to facilitate the biomass acclimatization to the substrate under investigation³⁷.

Results and discussion

Mono substrate digestion

Biodegradability tests

The average values of specific methane production (SMP) obtained from the mono-substrate digestion were $0.346 \text{ L CH}_4 \text{ g}^{-1} \text{ ODM}$ for maize silage, and $0.326 \text{ L CH}_4 \text{ g}^{-1} \text{ ODM}$ for red *radicchio* waste, both resulting in agreement with literature data^{19,21,30}.

The conversion efficiency was evaluated by dividing the average actual SMP of each substrate for its theoretical methane potential $B_{0, \text{th ODM}}$, obtaining good performances for both feeds: 81.4 % for maize silage, and 78.4 % for red *radicchio* waste. It has to be pointed out that the practical methane yield is always lower than the theoretical one, due to several factors, such as: the utilization of a substrate fraction to synthesize new bacterial mass, the finite retention time, the limitation of nutrient factors, and the usual inaccessibility of a part of particulate organic substrates³³. Undoubtedly, lignin plays a role in hampering the degradation of cell wall constituents. Herrmann *et al.*²² found a negative correlation between the specific methane production of crop silage and the parameters that describe the fiber

fraction (namely, NDF and ADL). In order to enhance the anaerobic digestion of lignocellulosic matrices, Planinic *et al.*²³ developed a pretreatment system of the corn forage by the white-rot fungus *T. versicolor*, obtaining a lignin degradation of 71 %. Applying the aforementioned pretreatment to corn silage before the co-digestion with cow manure, Tisma *et al.*³⁸ noticed an increase in methane generation rate.

The lower conversion efficiency of red chicory could also be due to its low value of C:N ratio. Astals *et al.*¹⁵ pointed out that values of C:N from 20 to 60 are reported in literature as optimum for anaerobic digestion. Nevertheless, other authors observed long-term stable operation in a two-stage process (solid-state anaerobic digestion + granular biomass reactor) treating brewery spent grain at a very low C:N ratio ranging from 0.16 to 4.69³⁹.

Solubilization rate

Several authors agree that, for highly particulate organic matter, the initial hydrolysis is the rate-limiting step of the entire anaerobic digestion process^{33,37,40}. In ADM1, the breakdown of complex organic substrates to soluble monomers is modeled through the disintegration and hydrolysis stages⁴¹: during disintegration, the composite substrate is divided into particulate carbohydrates, proteins and lipids that are further degraded, by the hydrolysis stage, into monosaccharides, amino acids and long-chain fatty acids. The disintegration step and the three hydrolysis processes (for carbohydrates, proteins, and lipids, respectively) are assumed to follow first-order kinetics⁴¹.

The disintegration rate value (k_{dis}) suggested by ADM1 for mesophilic digestion is 0.5 d^{-1} , whereas the recommended hydrolysis constants for the carbohydrates, proteins, and lipids (k_{hyd_ch} , k_{hyd_pr} , k_{hyd_li}) are set at the same value of 10 d^{-1} .

In the present study, the first-order rate k_{sol} (d^{-1}) was evaluated to describe the initial solubilization (disintegration + hydrolysis) of the complex organic substrates. As reported by Astals *et al.*⁴², the solubilization rate has been typically estimated from the cumulative methane production curve of the anaerobic biodegradability test, and then implemented as disintegration rate in ADM1, while a default non-limiting value is given to the carbohydrates, protein, and lipids hydrolysis rate. This is in agreement with Feng *et al.*⁴³, suggesting that the high values for k_{hyd_ch} , k_{hyd_pr} and k_{hyd_li} are proposed in order to completely exclude the influence of hydrolysis step in the model.

Assuming that methane production was mainly limited by the solubilization rate with no accumulation of intermediate products, the cumulative meth-

Table 3 – Evaluation of solubilization parameters

Parameter	Units	MS	RR
SMP_{ult}	[STP L CH ₄ g ⁻¹ ODM]	0.356	0.331
k_{sol}	[d ⁻¹]	0.231	0.389

Table 4 – Blend characteristics

Parameter	Unit	Average value
TS	[%]	21.6
VS	[% DM]	91.1
Carbohydrates (*)	[% DM]	73.3
Lignin	[% DM]	4.1
Proteins	[% DM]	13.0
Lipids	[% DM]	2.1
C:N	–	26
B_{0th}	[STP L CH ₄ g ⁻¹ ODM]	0.421

ane production was represented by a first-order kinetic for the solubilization of particulate organic matter⁴⁰:

$$SMP_{(t)} = SMP_{ult} \cdot \left(1 - e^{(-k_{sol} \cdot t)}\right) \quad (4)$$

where $SMP_{(t)}$ is the specific methane production (L CH₄ g⁻¹ ODM) at time t at standard conditions (STP), SMP_{ult} is the ultimate methane potential (*i.e.*, the production achievable at infinite residence time), and k_{sol} is the total disintegration/hydrolysis constant.

The values of SMP_{ult} and k_{sol} (reported in Table 3) were evaluated using non-linear least squares curve fitting on the net cumulative specific methane production (Fig. 2).

Co-digestion tests

The co-digestion tests were carried out by mixing maize silage and red chicory waste with a

MS:RR (in ODM basis) ratio ranging from 1.3 to 1.5 and a substrate to inoculum ODM ratio of 8 % – 9 %. The blend composition (reported in Table 4) assured the optimal value of 26 for the C:N ratio.

The theoretical methane potential of the co-substrates (also reported in Table 4) was predicted assuming the additivity of methane production obtaining a value of $B_{0,th ODM}$ of 0.421 L CH₄ g⁻¹ ODM (in STP conditions).

The measured value of SMP was 0.371 (STP) L CH₄ g⁻¹ ODM with a conversion efficiency of 88.1 %.

To investigate the effects of co-digestion, an initial hypothesis of no-synergism was done so the blend methane production was simulated according to equation (4), modified to take into account the mixing ratio:

$$SMP_{blend(t)} = x_{MS} \cdot SMP_{ult MS} \cdot \left(1 - e^{(-k_{sol MS} \cdot t)}\right) + x_{RR} \cdot SMP_{ult RR} \cdot \left(1 - e^{(-k_{sol RR} \cdot t)}\right) \quad (5)$$

where x_{MS} and x_{RR} represent the ODM fraction of single substrate with respect to the blend feed.

In Fig. 3, the comparison between the measured SMP (scattered curve) and the predicted one (continuous line) calculated with equation (5) is shown. As it is evident, the measured production curves are always above the calculated ones, in agreement with Mata-Alvarez *et al.*⁴⁴ and Aichinger *et al.*⁴⁵ stating that, under favorable conditions, 1+1>2 may be achieved, *i.e.*, the co-digestion of two substrates can produce more methane than the addition of the methane produced in both single-substrate digestions. The synergism between the two substrates had a clear beneficial effect on the solubilization rate: as shown in Fig. 3, the actual disintegration/hydrolysis step was faster than the predicted one.

In order to measure the observed synergistic effects, two parameters were considered: ΔSMP_{blend} and Δk_{blend} .

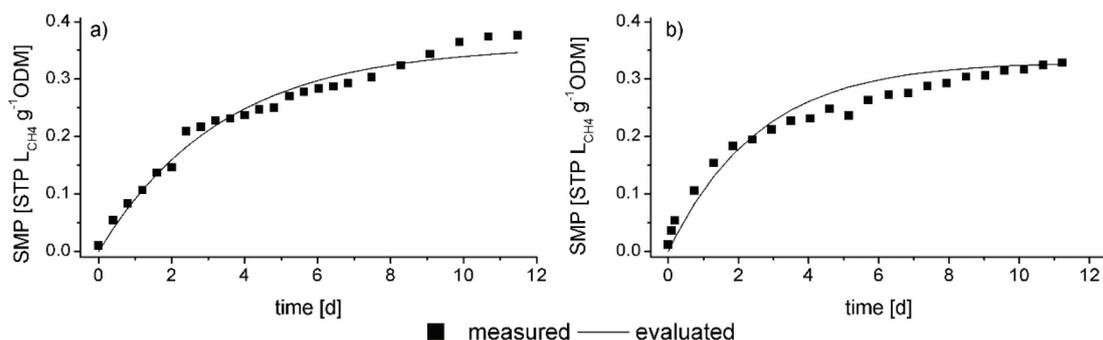


Fig. 2 – Fitting results: a) MS, b) RR

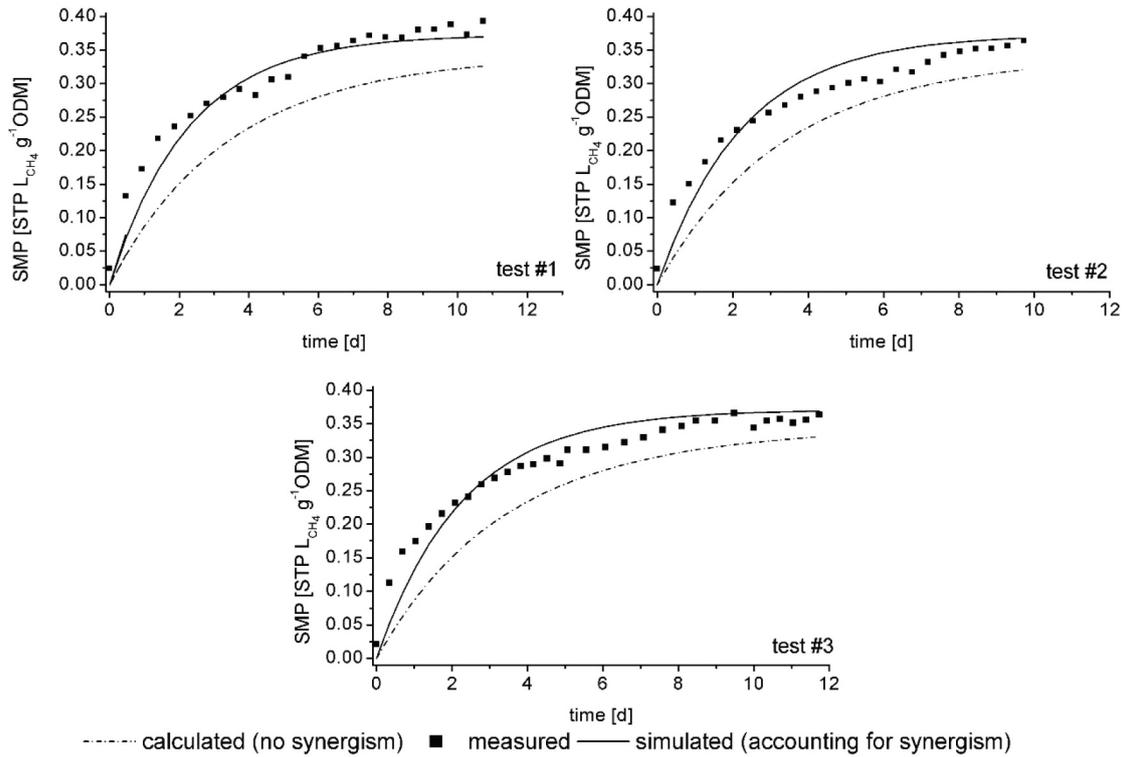


Fig. 3 – Evaluation of synergistic effects: scattered curve represents the measured methane production, dotted curve represents the predicted production in the hypothesis of no synergism, continuous line is the simulated production accounting for synergism

The ΔSMP_{blend} represent the relative change in SMP_{ult} compared to what would be expected based on mono-digestion tests⁴⁶. Its value was determined as:

$$\Delta SMP_{blend} = \frac{SMP_{blend,ult}}{(SMP_{MS,ult} \cdot x_{MS} + SMP_{RR,ult} \cdot x_{RR})} - 1 \quad (6)$$

If co-digestion had no effect on methane yield compared to mono-digestion, ΔSMP_{blend} would equal 0.

Analogously, Δk_{blend} is the relative change in solubilization rate constant due to co-digestion:

$$\Delta k_{blend} = \frac{k_{blend}}{(k_{MS,sol} \cdot x_{MS} + k_{RR,sol} \cdot x_{RR})} - 1 \quad (7)$$

The kinetic equation (4) was then reformulated as:

$$SMP_{(t)} = SMP_{blend} \cdot \left(1 - e^{(-k_{sol,blend} \cdot t)}\right) \quad (8)$$

where SMP_{blend} and $k_{sol,blend}$ are, respectively:

$$SMP_{blend} = (SMP_{MS,ult} \cdot x_{MS} + SMP_{RR,ult} \cdot x_{RR}) \cdot (1 + \Delta SMP_{blend}) \quad (9)$$

$$k_{blend} = (k_{MS,sol} \cdot x_{MS} + k_{RR,sol} \cdot x_{RR}) \cdot (1 + \Delta k_{sol,blend}) \quad (10)$$

Table 5 – Co-digestion parameters

Parameter	Unit	Value
$SMP_{blend,ult}$	[STP L CH ₄ g ⁻¹ ODM]	0.373
$k_{sol,blend}$	[d ⁻¹]	0.439
ΔSMP_{blend}	[-]	0.08
$\Delta k_{sol,blend}$	[-]	0.48

Fig. 3 shows the specific methane production simulated with equation (8). As highlighted by simulation results, presented in Table 5, the synergism due to the co-digestion had a marked effect mainly on the solubilization rate, the value of which was higher than that obtained with mono-digestion of MS and RR. Indeed, the relative increment Δk_{blend} was found significantly higher than zero. Lower improvements were observed in the ultimate methane potential, with a relative change ΔSMP_{blend} equal to 0.08. Similar results were found by Astals *et al.*¹⁵ who concluded that AcoD leads to an improvement mainly of the AD kinetics.

Conclusions

The synergistic effects due to anaerobic co-digestion of maize silage and red chicory waste were assessed by means of anaerobic biodegradability

tests followed by methane production simulation. Results showed that blending the two substrates led to an improvement of the digestion process in comparison with the performance obtained from the mono-digestion of each single substrate. Synergistic effects concerned the solubilization process of the complex particulate substances with no significant increase in the ultimate methane potential.

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